

# Seasonal variation in the response of arbuscular mycorrhizal fungi to grazing intensity

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## Abstract

Despite existing evidence of pronounced seasonality in arbuscular mycorrhizal (AM) fungal communities, little is known about the ecology of AM fungi in response to grazing intensity in different seasons. Here, we assessed AM fungal abundance, represented by soil hyphal length density (HLD), mycorrhizal root colonisation intensity (MI) and arbuscule intensity (AI) throughout three seasons (spring, summer, autumn) in a farm-scale field experiment in typical, grazed steppe vegetation in northern China. Seven levels of field-manipulated, grazing intensities had been maintained for over 13 years within two topographies, flat and slope. We also measured soil nutrients and carbon content throughout the growing season to investigate whether seasonal variation in AM fungal abundance was related to seasonal shifts in soil resource availability along the grazing gradient. We further examined the association between AM fungal metrics in the different grazing treatments through the growing season. Our results showed a pronounced seasonal shift in HLD but there was no clear seasonality in MI and AI. HLD was significantly negatively related to grazing intensity over the course of the growing season from spring to autumn. However, MI and AI were related negatively to grazing intensity only in spring. In addition, differential

responses of AM fungal abundance to grazing intensity at the two topographical sites were detected. No strong evidence was found for associations between AM fungal abundance and soil resource availability. Moreover, AM fungal internal and external abundance were correlated positively under the different grazing intensities throughout the growing season. Overall, our study suggests that external AM fungal structures in soil were more responsive to seasonal variation and grazing than internal structures in roots. The findings also suggest that early grazing may be detrimental to AM fungal root colonization of newly-emerged plants.

**Keywords:** spring grazing, topography, soil resource availability, grazing management, external hyphae, seasonal variation

## Introduction

Grasslands play a crucial role in global ecosystem functioning and human well-being (O'Mara 2012; Steinfeld et al. 2006). However, many grasslands are currently facing many pressures, of which overgrazing is one of the key drivers reducing grassland productivity and sustainability (Conant 2010; O'Mara 2012). Pervasive excessive grazing has altered above-ground plant communities, soil water and nutrient availability in grasslands (Conant 2010; McSherry and Ritchie 2013). This could translate to changes in the intimately-connected below-ground microbial community including the most common symbionts in grasslands, arbuscular mycorrhizal (AM) fungi (Birgander et al. 2014; Regan et al. 2014).

AM fungi are keystone soil micro-organisms that play a vital role in maintaining grassland ecosystem productivity and stability (Asmelash et al. 2016; Moora and Zobel 2010). Root symbiotic mycorrhizal fungi establish these mutualistic symbioses with a large proportion of terrestrial plant taxa (over 80%) (Brundrett and Tedersoo 2018). This association is fundamentally a nutritional symbiosis: AM fungi rely on photosynthetic carbon received from the plant in exchange for transfer of nutrients, in particular

phosphorus (Doubková et al. 2013; Zavalloni et al. 2012). As such, AM fungi can enhance plant grazing-tolerance by improving nutritional status, and thereby improving plant productivity (Moora and Zobel 2010; Walling and Zabinski 2006).

On the other hand, long-term grazing can alter AM fungal function and communities (Ba et al. 2012; Guo et al. 2016). The effect of grazing on mycorrhizal fungi can be explained in part by the carbon limitation hypothesis; clipping and removing plant photosynthetic tissues through long-term grazing may cause a decrease in carbon allocation to roots and mycorrhizal fungi as a result of competition between the plant and AM fungi for limited carbon resources (Gehring and Whitham 2003). Therefore, a negative response of AM fungi to long-term herbivory is expected. However, contradictory results have been reported (Barto and Rillig 2010; Faghihinia et al. 2020), so research is needed to improve our predictions of grazing effects on ubiquitous symbiotic AM fungi.

The extent of the grazing impact on AM fungal function and community structure depends largely on grazing intensity (Ba et al. 2012; Yang et al. 2020) as it has disparate impacts on above- and below-ground productivity and biodiversity (Yan et al. 2013). Whilst overgrazing has destructive and irreversible negative impacts on plant community and soil properties, under-grazing can be just as harmful as overgrazing to grassland biodiversity and functioning through less stimulation of plant growth and loss of grazing-dependent legumes and grasses (Metera et al. 2010). However, under-grazing is not a common practice worldwide at the moment. On the other hand, moderate grazing has been indicated as a benefit to grassland plant and soil conditions through natural fertilization, seed dispersal, making room for annual and bi-annual plant species growth and expansion, and periodic above-ground defoliation which regulate succession in plant communities (Metera et al. 2010). However, the effects of different grazing intensities on AM fungi have not been sufficiently addressed (van der Heyde et al. 2019).

In addition, the impact of grazing intensity on AM fungal structures may not be significant at particular time points throughout the growing season (Faghihinia et al. 2020). Many studies address the response of AM fungi to grazing at a single seasonal time point (Bai et al. 2013; van der Heyde et al. 2017), with few assessing the seasonal shift in AM fungal responses to herbivory, particularly in temperate systems (Cavagnaro et al. 2019; Staddon et al. 2003b; Wang et al. 2014). Cavagnaro et al. (2019) showed that AM fungal root colonization was significantly greater in summer compared with autumn in both sheep-preferred and non-preferred plant species in a steppe grassland, Argentina. Similarly, Staddon et al. (2003b) and Wang et al. (2014) reported greater mycorrhizal root colonization in the summer but lower values during the autumn in temperate ecosystems. The same seasonal pattern has been reported for fungal hyphal length density in soil (Staddon et al. 2003b). Summer peaks in AM fungal abundance are expected due to greater plant mineral nutrient demand, rapid vegetative growth and root production as a result of high light availability for photosynthesis. In addition, plants may need additional mineral nutrients to fund shoot regrowth, thus allocating more carbon to AM fungi during summer when grazing is most intense (Cavagnaro et al. 2019).

Moreover, the temporal dynamics of AM fungal abundance are confounded by ecosystem complexity, often with no consistent pattern being reported. For instance, no seasonal variation in AM fungal root colonization between summer and winter was reported in a Danish coastal, sandy, temperate grassland (Lekberg et al. 2013). However, another study of seasonal dynamics of AM fungal abundance in five Mediterranean plant species found that the percentage mycorrhizal root colonization and density of external hyphae was greater in autumn than in spring (Varela-Cervero et al. 2016). Further research is, therefore, required to unravel the underlying mechanisms of seasonality impact on AM fungal function and community which has prominent implications for grassland ecosystem management and stability.

Pronounced seasonality in AM fungal abundance also is likely to be attributed to soil resource availability which changes seasonally (Hewins et al. 2015; Wang et al. 2014). Hewins et al. (2015) showed that plant

nitrogen and phosphorus content increased from late summer to early spring and the observed trend was associated with a decline in mycorrhizal root colonization in a forest herb in northeastern Ohio, USA. Wang et al. (2014) found a positive correlation between temporal changes of AM fungal root colonization and spore richness as well as soil acid phosphatase activity and available phosphorous in temperate grasslands in the north of China. Seasonal shift in below-ground carbon allocation to AM fungal storage lipids (16:1 $\omega$ 5 NLFA) was observed in a coastal grassland in Denmark (Lekberg et al. 2013). In addition, soil resource availability alters along topographic gradients through topographical-induced changes in soil moisture and nutrient availability and solar exposure (Faghihinia et al. 2020; Murray et al. 2010; Schowalter 2016). Topography also affects animal behavior and distribution via greater livestock density and larger loads of dung and urine in low-lying areas compared with areas at higher elevation (Johnson et al. 2016). Topographic gradients of moisture and nutrient availability may interact with grazing to influence AM fungal variables, but the interaction under natural environments has yet to be discerned. How seasonal variation in AM fungal abundance relates to topographic-induced change in soil resource availability also requires further exploration.

An additional consideration is that AM fungi inhabit two different environments, inside host plant roots and in the surrounding soil. Given that these two media differ in terms of AM fungal community structure (Li et al. 2018; Stevens et al. 2020) and are exposed to disparate biotic interactions (Jansa et al. 2013), various responses of AM fungal internal and external structures to environmental disturbance is highly likely. The external hyphal network in soil has a shorter life-span and higher turnover than internal hyphae within the roots (Varma and Hock 2013), thus, external hyphae respond very quickly to seasonal environmental variations such as pulses in soil moisture and nutrient availability (Treseder et al. 2010). Yet, whether any association exists between AM fungal root colonization and external hyphae in response to grazing intensity over growing seasons is unclear. Examining the linkage between AM fungal abundance

in roots and soil is fundamental for some of the crucial functional features of plant-fungal symbiosis including plant nutrient acquisition from the soil and host plant productivity (Jansa et al. 2013).

As far as we are aware, no information is available concerning the interaction of grazing intensity, season and topography on AM fungal abundance (represented by soil hyphal length density (HLD), mycorrhizal root colonization intensity (MI) and arbuscule intensity (AI)). We aimed, therefore, to answer the following three questions (1) Is there a temporal change in mycorrhizal abundance in response to grazing intensity?, (2) Does seasonal variation in AM fungal abundance relate to seasonal shifts in soil resource availability along a grazing gradient in two topographic locations? (3) Is there any association between AM fungal abundance in soil and roots throughout the growing season?

## **Methods**

### **Study Site**

This study was conducted at the Sino-German grazing experimental site in the Xilin River Basin of Inner Mongolia, China (116° 42' E; 43° 38' N), a steppe grassland ecosystem with a semi-arid, continental climate. We set up our experiment in 14 plots located in two topographic blocks, flat and slope. The “slope block” had a topographical slope of about 8 degrees, and the “flat block” had no noticeable slope. The two topographic blocks, were significantly distinct in terms of soil moisture, soil bulk density and soil nutrient availability as well as plant community structure and species aggregation (Li et al. 2017; Li et al. 2015; Ren et al. 2018). Each experimental plot, encompassing an area of 2 ha, was subjected to one of seven levels of grazing intensity (GI), from 0 to 9 ewes per ha with an interval increase of 1.5 ewes (35 kg live-weight female sheep). Hereafter, we represent GI by the number of grazers per hectare as 0 (no grazing), 1.5 (very light), 3 (light), 4.5 (light-moderate), 6 (moderate), 7.5 (heavy) and 9 (overgrazing). Ewes were put

in plots for 90 days during the grazing season from June to September each year. Until we took samples in 2018, the grazing experiment had been run continuously for 13 years. Plant communities of both topographies are dominated by two perennial C3 grasses, *Leymus chinensis* (Trin.) Tzvel. a rhizomatous grass, and *Stipa grandis* P. Smirn. a bunchgrass, which together account for more than 75% of the total above-ground biomass (Li et al. 2017). A detailed description of the climate, vegetation cover, soil characteristics and the design of the experimental site can be found in previously published papers (Schönbach et al. 2011; Wan et al. 2011) and in the supplementary information (SI-1).

#### **Soil sampling**

Five evenly-distributed double soil core samples (2 × 20 cm) were collected from each plot over the growing season at three sampling times in 2018 (2 topographical locations × 7 levels of grazing intensity × 3 seasons × 5 samples). In total 210 soil core samples for mycorrhizal measurement and 210 for soil properties analyses were collected. In the study area, the growing season begins in May, peaks in July and ends in September while the grazing season starts from June continuously to the end of September (Wan et al. 2011). We took samples in early-May, when grazing had not been started yet, mid-July, in the middle of the grazing season and late-September, at the end of the grazing season, representing spring, summer and autumn collections, respectively. Soil samples were kept in an ice box with a temperature of around 0 °C until being placed in storage at -20°C. A schematic illustration of the experimental design is presented in Appendix SI-3 Figure S1.

#### **Soil hyphal length density (HLD) measurement**

Soil hyphae were extracted from two sub-samples of 5 g soil from each soil core (420 samples in total) in 500 ml of deionized water (dH<sub>2</sub>O) following a modified membrane filter technique from Jakobsen et al. (1992) and Boddington et al. (1999). The hyphae of AM fungi were identified based on microscopic features; angular, aseptate, and 1.0–13.4 µm in diameter (Boddington et al. 1999; Shen et al. 2016). The

total length of hyphae (mm) was measured for a minimum 60 fields of view for each filter at  $\times 100$  magnification. A modified GIM (Gridline Intersect Method) equation based on (Tennant 1975) was used for calculating the total length of hyphae (mm) per gram of soil ( $\text{m g}^{-1}$ ) (Shen et al. 2016) (SI-2).

#### **Mycorrhizal root colonization assessment**

Roots, comprising multiple plant species, were collected from five soil cores from each plot. The roots were rinsed carefully with distilled water and a sonicator was used to remove soil particles adhering to the root surface. Roots were cut into pieces *ca.* 1 cm long and then approximately 5 g of fine roots of each sample were cleared in 2% KOH (w/v) at 90°C for 60 min and then rinsed thoroughly on a fine sieve before being acidified in 2% HCl (v/v) for 30 min and stained in 0.05% (w/v) trypan blue: glycerol: lactic acid (1:2:1) for 30 min at 90 °C. Root segments of each sub-sample were rinsed with lactic acid: glycerol: dH<sub>2</sub>O (1:2:1), selected randomly and mounted on slides in 50% glycerol. Thirty pieces of roots from each root sub-sample were observed under the compound microscope (Nikon eclipse Ci-L) at  $\times 200$  and  $\times 400$  magnification, and mycorrhizal colonization intensity in the root system (MI%) (Percentage of total segment length colonized) and arbuscule intensity (AI%) (arbuscular abundance in the root system) were assessed according to the five-class system of Trouvelot (1986). Although assessed, mycorrhizal frequency was uniformly high and was not informative (data not shown).

#### **Soil resource availability determination**

Fresh soil samples were air-dried and sieved through a 2-mm sieve. Soil organic carbon was determined by the acid-potassium dichromate oxidation method (Walkley and Black 1934). Soil available phosphorus (Olsen-P) was extracted with NaHCO<sub>3</sub> and determined by spectrophotometry following (Olsen 1954) and soil available nitrogen was measured by the alkali-hydrolyzed diffusion method according to Bao (2000).

#### **Data analysis**



We used a three-way nested design to test the interactive effects of grazing intensity, topography and season on AM fungal measures. The data are nested in the sense that sampling was conducted at two sites with contrasting topography; flat and slope. At each topographic location, samples were collected from seven plots, each with different levels of grazing intensity, and sampling was repeated in three seasons (Appendix SI-3, Figure SI1).

We conducted three analyses on our nested hierarchical data. First, we assessed grazing, topography and season effects and their interaction on AM fungal variables by linear mixed effect models (LMEs). Response variables included (i) soil hyphal length density (ii) mycorrhizal root colonization intensity and (iii) arbuscule intensity. Explanatory variables were grazing intensity with interaction with season, and study plot (nested by topography and grazing intensity) was a random variable. LME models fitted by maximum likelihood were applied separately for each AM fungal response variable. Due to the design of this large scale long-term field experiment, we treated grazing intensity as a continuous variable. We first fitted a model with all terms as well as all their interactions. Then, automated model selection using Akaike's information criterion (AIC) were carried out to find the best-fit model.

Second, we assessed the relationship between AM fungal hyphal length density, mycorrhizal root intensity and arbuscule intensity and soil resource variability including (i) available nitrogen, (ii) organic carbon and (iii) available phosphorus in topographic sites using linear regression models. As the effect of environmental conditions on AM fungal responses might not be independent within our soil cores, but could be homogeneous within each plot, we pooled data from the same plot, and analyzed the relationship between AM fungal measures and the means of environmental variables for plots. According to Crawley (2012) and Zuur et al. (2009) statistical analysis on nested data with hierarchical structure should be carried out on means rather than on individual observations so as to provide a conservative estimate of significance and to reduce the likelihood of Type I errors.

Third, we examined the responses of soil resource variables to different grazing intensities by linear mixed effect models. Linear mixed effect models were applied separately to soil available nitrogen, phosphorus and organic carbon, and study plot (nested by topography and grazing intensity) was a random variable. The best-fit models were then selected based on AIC. Finally, to assess relationships between HLD and the other AM fungal variables, Pearson correlation coefficients ( $r$ ) among the means per plot were calculated. All data analyses were conducted with R, version 3.6.2 (R Core Team 2018). Linear mixed effect models were applied using the `lme()` function from the “nlme” package (Pinheiro et al. 2018). Automated model selections were carried out with the package “MuMIn” using the ‘dredge’ function (Barton 2018) to find the best-fit models and statistical inference. All models were validated by checking the distribution of residuals following Zuur et al. (2009). Visual inspection of residual plots did not reveal any noticeable deviations from normality or homoscedasticity.

## Results

### Seasonality and grazing intensity effects on AM fungal abundance at two topographies

A strong negative relationship between grazing intensity and soil hyphal length density (m/g) (HLD) was detected in all three seasons. HLD decreased significantly with increasing grazing intensity, and this was evident in all seasons: spring ( $\beta = -0.23 \pm 0.06$ ,  $P = 0.002$ ), summer ( $\beta = -0.27 \pm 0.06$ ,  $P = 0.001$ ) and autumn ( $\beta = -0.43 \pm 0.06$ ,  $P < 0.001$ ) in the flat area, as well as in spring ( $\beta = -0.48 \pm 0.06$ ,  $P < 0.001$ ), summer ( $\beta = -0.45 \pm 0.06$ ,  $P < 0.001$ ) and autumn ( $\beta = -0.54 \pm 0.06$ ,  $P < 0.001$ ) in the slope area (Figure1, Table 1).

HLD did vary during the growing seasons but the trends were different in the two topographic locations. HLD increased in the flat area, but decreased in the slope area over the course of the growing seasons. HLD significantly increased from spring to summer ( $\beta = 1.70 \pm 0.45$ ,  $P < 0.001$ ) and from summer to autumn

230 ( $\beta=2.13\pm0.45$ ,  $P < 0.001$ ) in the flat area. In contrast, HLD decreased from spring to summer ( $\beta=-1.89\pm0.45$ ,  
 231  $P < 0.001$ ) with no significant difference between summer and autumn in the slope area ( $\beta=-0.03\pm0.45$ ,  
 232  $P=0.952$ ) (Figure 2). Significantly higher HLDs were observed in the slope site in spring ( $\beta=-4.09\pm0.45$ ,  $P <$   
 233  $0.001$ ) and autumn ( $\beta=1.66\pm0.45$ ,  $P < 0.001$ ) compared with the flat site (Table 1).

234 There was a negative relationship between mycorrhizal root intensity (MI) and grazing intensity in spring  
 235 but no such relationship in summer and autumn. As grazing intensity increased, MI decreased marginally  
 236 in spring ( $\beta=-0.91\pm0.43$ ,  $P=0.057$ ) in the flat area, and significantly in spring ( $\beta=-2.02\pm0.43$ ,  $P < 0.001$ ) in  
 237 the slope area (Figure1, Table 1).

238 MI variation along the growing season differed at the two topographic locations: MI increased in the flat  
 239 area while it decreased in slope area during the growing seasons (Figure 2). No significant difference was  
 240 observed between spring and summer ( $\beta=-6.43\pm3.29$ ,  $P=0.052$ ) but MI significantly increased from  
 241 summer to autumn ( $\beta=9.16\pm3.29$ ,  $P=0.006$ ) in the flat area. In contrast, MI significantly decreased from  
 242 spring to summer ( $\beta=-11.21\pm3.29$ ,  $P=0.001$ ) and remained unchanged from summer to autumn ( $\beta=-$   
 243  $1.60\pm3.29$ ,  $P=0.627$ ) in the slope area (Figure 2, Table 1). Topography significantly impacted MI with higher  
 244 abundance in slope site in spring ( $\beta=-9.57\pm3.29$ ,  $P= 0.004$ ) but not in summer ( $\beta=-4.79\pm3.29$ ,  $P=0.1473$ )  
 245 and autumn ( $5.97\pm3.29$ ,  $P=0.072$ ) (Table 1).

246 Arbuscule intensity (AI) showed the same pattern as MI to grazing intensity with a negative relationship  
 247 with grazing intensity in spring in the flat ( $\beta=-0.42\pm 0.15$ ,  $P=0.017$ ) and slope sites ( $\beta=-0.56\pm0.15$ ,  $P=0.003$ )  
 248 (Figure 1, Table 1). No grazing intensity effect on AI was found in summer and autumn. No seasonal shifts  
 249 were found in AI in the flat area, while AI significantly decreased from summer to autumn ( $\beta=-3.88\pm1.14$ ,  
 250  $P=0.001$ ) in the slope area (Figure 2, Table 1). AI was significantly higher in the slope site throughout the  
 251 growing season in spring ( $\beta=0.42\pm0.15$ ,  $P=0.017$ ), summer ( $\beta=-2.62\pm1.14$ ,  $P=0.023$ ) and autumn  
 252 ( $\beta=2.48\pm1.14$ ,  $P=0.031$ ) compared with the flat site (Table 1).

The interaction of grazing intensity and topography was only significant for HLD but not for MI and AI. Model fitting and selection revealed non-significant effects of three way interactions between grazing intensity, topography and season on HLD, MI and AI (Table 1).

#### **Seasonality and grazing intensity effects on soil resource availability at two topographies**

Soil available nitrogen (AN) (mg/kg) was not related to grazing intensity at the two topographic locations but it did change over the growing season in the flat area; AN increased from spring to summer ( $\beta=10.28\pm3.38$ ,  $P=0.003$ ) and then decreased from summer to autumn ( $\beta=-8.21\pm3.38$ ,  $P=0.016$ ). AN marginally increased from spring to summer ( $\beta=5.99\pm3.38$ ,  $P=0.079$ ) while it remained unchanged from summer to autumn in the slope site (Table 1, Figure S3). The flat site exhibited significantly higher availability of soil nitrogen in summer ( $\beta=-10.52\pm3.38$ ,  $P=0.002$ ) and autumn ( $\beta=-7.63\pm3.38$ ,  $P=0.025$ ) compared with the slope site (Table 1, Figure S3).

Soil available phosphorus (AP) (mg/kg) was related positively to grazing intensity in spring in the flat area ( $\beta=0.31\pm0.08$ ,  $P=0.004$ ) and in autumn in the slope area ( $\beta=0.43\pm0.08$ ,  $P < 0.001$ ) (Table 1, Figure S2). AP did not change along the growing season in the flat area but it significantly, though marginally, decreased from summer to autumn in the slope area ( $\beta=-1.32\pm0.64$ ,  $P=0.042$ ) (Table 1, Figure S3). There was significantly greater phosphorus availability in summer ( $\beta=-1.79\pm0.64$ ,  $P=0.006$ ) and autumn ( $\beta=-3.43\pm0.64$ ,  $P < 0.001$ ) in the flat site compared to the slope site (Table 1, Figure S3).

Soil organic carbon (SOC) (%) was related negatively to grazing in spring ( $\beta=-0.06\pm0.03$ ,  $P=0.052$ ) and autumn ( $\beta=-0.07\pm0.03$ ,  $P=0.033$ ) in the flat area and in summer ( $\beta=-0.1\pm0.03$ ,  $P=0.005$ ) in the slope area (Table 1, Figure S2). Pronounced seasonality was observed for SOC but the pattern differed between the topographic locations. SOC decreased from spring to summer ( $\beta=-0.69\pm0.22$ ,  $P=0.002$ ) and then increased in autumn ( $\beta=0.49\pm0.22$ ,  $P=0.027$ ) in the flat area. In contrast, SOC significantly increased from spring to

summer ( $\beta=0.60\pm0.22$ ,  $P=0.007$ ) and decreased from summer to autumn ( $\beta=-0.78\pm0.22$ ,  $P=0.001$ ) in the slope area (Table 1, Figure S3).

#### **Relationship between AM fungal abundance and soil resource availability**

HLD was not related to any measured variables (Table S1, Figure S4 and S5). MI was related negatively to soil available nitrogen in the flat site ( $\beta=-0.36\pm0.15$ ,  $P=0.029$ ) and AI was significantly negatively related to soil available phosphorus ( $\beta=-0.84\pm0.29$ ,  $P=0.01$ ) in the flat site (Table S1, Figure S4 and S5).

#### **Relationship between AM fungal abundance in soil and roots**

There was a significant positive association between HLD and MI in both the flat (Pearson  $r=0.49$ ,  $P=0.024$ ) and the slope sites (Pearson  $r=0.61$ ,  $P=0.003$ ) throughout the growing season (Figure 3). HLD was significantly correlated with AI in flat site (Pearson  $r=0.54$ ,  $P=0.011$ ), but no significant association was detected for the slope site (Pearson  $r=0.37$ ,  $P=0.103$ ).

#### **Discussion**

Climatic seasonality and inter-annual variations in temperature and precipitation are expected to moderate the effects of grazing on plant and soil related factors, and thereafter on below-ground biota including mycorrhizal fungi. Nevertheless, the interaction of seasonality and grazing effects on AM fungal abundance has not been investigated fully, particularly for HLD (Faghihinia et al. 2020). Our findings demonstrated significant negative relationships between HLD and grazing intensity and this trend persisted in all three seasons (Figure 1). The negative response of HLD to grazing has been reported previously in several studies in grassland ecosystems (Ren et al. 2018; van der Heyde et al. 2017; Vowles et al. 2018). Grazing-induced reduction in above-ground vegetation cover and below-ground root biomass (Hao and He 2019) would reduce the range of plants root types and the range of root exudates (Wilson et al. 2018) which would consequently impact soil microbes including AM fungi (Wang et al. 2014). Given

that hyphal extension and germination of AM fungal spores is known to take place preferentially in the presence of roots and root exudates (Smith and Read 2008; Tahat et al. 2010), reduction in HLD with increasing grazing intensity is expected. Noteworthy, however, is the consistent trend in the response of HLD to long-term grazing intensity from early in the season to the end of the growing season supporting the hypothesis that the effects of grazing intensity on external hyphal abundance is moderated by seasonality.

Seasonal dynamics were not pronounced in mycorrhizal root colonization variables. MI and AI were significantly negatively related to grazing intensity only in spring but not in the summer and autumn (Figure 1). One possible explanation is that plants allocate less carbon to below-ground root colonizers in spring due to lack of mature leaf tissues and thus lower total photosynthetic activity (Hewins et al. 2015). Plants generally allocate more carbon to leaf elongation rather than root growth at the early stages of their growth (Waterton and Cleland 2016), suggesting that mycorrhizal root colonization is most likely governed by plant physiological status. Moreover, herbaceous vegetation is susceptible to herbivory during the early stage of the growing season due to small plant sizes, undeveloped physical (e.g., hard shells, thorns or spines) and low chemical defense mechanisms (e.g., producing secondary metabolites such as alkaloids, terpenoids, phenolics) as well as high palatability and nutritional quality (Quintero et al. 2014). This finding suggests that the potential impact of early grazing would not only be detrimental to newly emerged plants, as reported in previous studies (Quintero et al. 2014; Waterton and Cleland 2016), but also to AM fungal root colonizers. This impact on the mycorrhizal symbiosis has large implications for grassland management in term of the timing of grazing within the growing season. These insights can help with management decisions aimed at maintaining sustainable grassland productivity and soil biodiversity. Clear differential responses of AM fungal abundance to grazing intensity were observed between the two topographic locations. Overall, we observed higher HLD, MI and AI in the slope site compared with the flat site, particularly in the spring (Figure 2). Previous studies in the same site have shown that the two

topographies are distinct in terms of soil properties and plant communities. The flat site has a greater plant richness, above-ground biomass, soil nitrogen and phosphorus availability, soil moisture and pH compared with the slope site (Li et al. 2017; Ren et al. 2018; Schönbach et al. 2011; Wan et al. 2011). The slope area is therefore more nutrient-limited than the flat area. Given that the arbuscular mycorrhizal symbiosis involves a carbon and nutrient trade-off between the plant and fungal partners (Hodge et al. 2010), it is likely that plants are more dependent on mycorrhizal fungi for obtaining nutrients in the slope area and allocate more carbon below-ground in the search for additional nutrients (Johnson 2010). Plant demand for nutrients is greater in spring when they are in their rapid vegetative growth stage and leaf elongation takes place. There might not be as high a demand in the flat site at early stage of the growing season when soil mineral nutrients are abundantly available to plants compared with the more nutrient-limited slope area.

Furthermore, previous studies reported greater plant species richness (41 vs. 20 plant species) and above-ground biomass (129.02 vs. 77.06 g m<sup>-2</sup>) in the flat area compared to the slope area (Li et al. 2017; Wan et al. 2011). The heterogeneity of the plant community has resulted in a higher ecological threshold of community structure and ecosystem functioning to grazing intensity in the flat (3.75 sheep ha<sup>-1</sup>) compared with the slope site (3 sheep ha<sup>-1</sup>). As a result, it has been suggested that the plant community composition in the flat site is more resistant and resilient to grazing disturbance than that in the slope site (Li et al. 2017). Accordingly, AM fungi appeared more tolerant to some perturbations associated with grazing intensity in the flat site compared with the slope site because the corresponding plant community is itself more resilient to grazing impacts. This may in part explain the increasing HLD and MI from spring, when no grazing happens, to the end of growing season when grazing intensity is becoming intense. In contrast, reduction in HLD and MI throughout the growing season in the slope site was linked to lower nutrient availability and concomitant lower capability of plant species to respond to grazing pressure and defoliation.

The differential responses of AM fungal abundance at the two topographies can also be explained by differences among plant community composition and grazers' diet preferences. The vegetation at our experimental site is dominated by *L.chinensis* and *S.grandis*. The above-ground biomass and richness of *L.chinensis* is greater than that of *S.grandis* in the flat site (Schönbach et al., 2011; Wan et al., 2011). It has been shown that the above-ground biomass of palatable and highly mycorrhizal *L.chinensis* decreased substantially with increasing grazing intensity, whereas the biomass of relatively unpalatable, less mycorrhizal *S.grandis* remained unchanged along the grazing gradient (Wan et al. 2011). Thus, lower AM fungal abundance in the flat site could have been caused by the strong negative effects of grazing on dominant, palatable *L.chinensis*.

Seasonal differences in AM fungal abundance have been shown to be driven by shifts in relative abundance of soil resource availability (Hewins et al. 2015; Lekberg et al. 2013). However, we did not find similar seasonal trends in AM fungi and soil resources, and the relationships are not particularly strong for these variables. We found a marginally-significant negative relationship between (1) MI and soil available nitrogen, and (2) AI and soil available phosphorus in the flat site. Whether the seasonal shifts in AM fungal abundance are directly associated with soil resource availability cannot be confirmed in this current study and requires further investigation.

We found a strong positive association between hyphal length density in soil and the intensity of root colonization (Figure 3) suggesting that changes in AM fungal internal abundance in roots are positively associated with those of external abundance in soil over the growing season. The positive correlation between HLD and MI in our grazed study sites is not surprising because thin, fragile runner hyphae can be easily disrupted by the activities of large herbivores leading to lower nutrient uptake by the associated plants and lower redistribution of recently fixed carbon through the soil, thereby reduced colonization capacity of the fungi (Gui et al. 2018; van der Heyde et al. 2017). Such a relationship between the various metrics of AM fungal dynamics rarely has been reported in previous studies, particularly in grassland



ecosystems. Considering that nutrient uptake and carbon use differ among and within AM fungal structures (Smith and Read 2008) and that AM fungal isolates differ in their rates of colonization and hyphal extension (Hart and Reader 2005), this relationship between AM fungal structures may change across different ecosystems. Studies are needed to unravel underlying mechanisms.

## Conclusions

In summary, we showed that the effects of grazing intensity on AM fungal abundance is mediated by both topography and seasonality in this Inner Mongolia grassland. While we acknowledge that true replicates of each individual plot at a given grazing intensity and topography would have added increased robustness to our conclusions, to repeat such a large scale experiment with multiple large plots, in this case a total of 14 plots of 2 hectares each, is extremely expensive and unrealistic. By careful application of appropriate statistical analyses, our results clearly showed that, in the study site, HLD was negatively related to grazing intensity over the course of the growing season and MI and AI were significantly negatively related with grazing intensity only at the early stage of the growing season at both topographic locations.

That seasonal shifts in mycorrhizal abundance were more pronounced in HLD, but not so marked in MI and AI, suggest that external AM fungal structures in soil are more responsive to seasonal variation than internal mycorrhizal structures in roots. This can be explained by the mycorrhizal hyphae in the soil experiencing a much broader range of environmental conditions than those within the relatively stable conditions within plant roots. Indeed, soil HLD, containing a large proportion of AM fungal hyphae with short longevity and high turnover rate (Staddon et al. 2003a) is more susceptible to environmental disturbance compared with AM fungal hyphae inside roots (Varma and Hock 2013). MI and AI were significantly negatively related to grazing intensity only in spring which suggests that mycorrhizal root colonization is driven by plant physiology rather than by sampling time *per se*. Furthermore, our data

suggest that early grazing can be detrimental to AM fungal root colonization of newly emerged plants. Overall, early-spring grazing should be avoided in Inner Mongolia grazed steppe to prevent damage of plant growth and thereby their root-associated symbiotic partners. This will lead to maintaining healthy plant communities and soil biota with sustainable function of the grassland ecosystem.

## **Acknowledgements:**

This study was finically supported by XJTLU Research Development Fund (RDF-15-02-13). We appreciate the kind cooperation of the Inner Mongolia Grassland Ecosystem Research Station of the Chinese Academy of Sciences for kindly providing research station facilities and a plant dataset. The authors acknowledge the kind help of Lijia Dong, Zirui Hu, Yuwei He, Ye Wang, Chuanyu Zhou, Hanyu Jian, Fanjing Liu and Shasha Zou for collecting data in the lab.

## **Supplementary data**

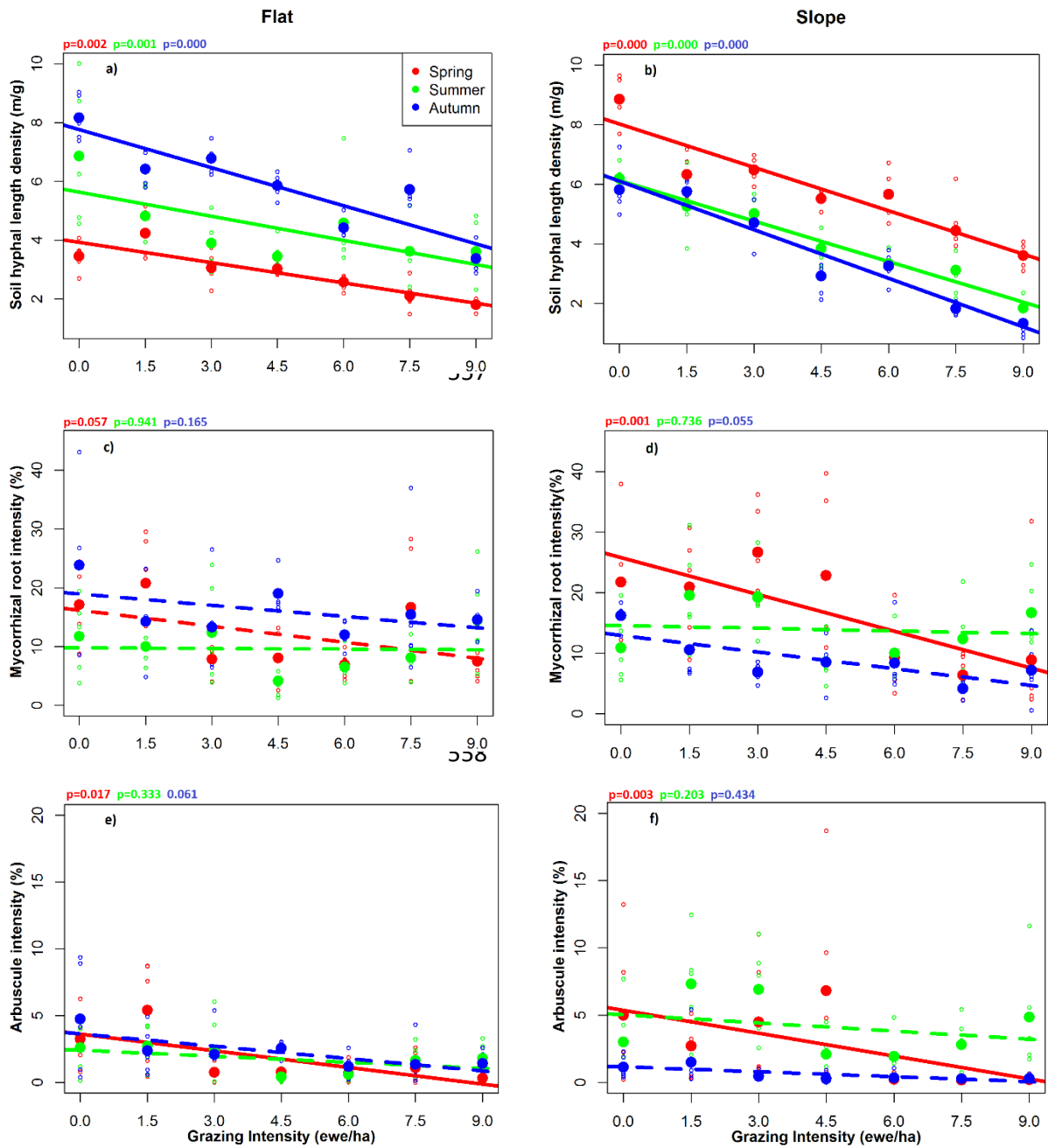
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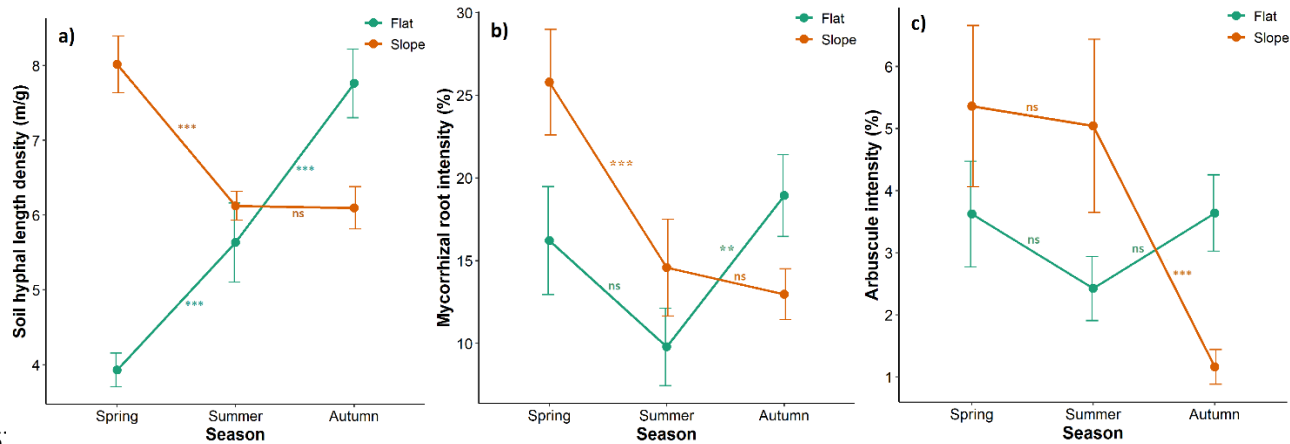
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**Figure 1.** Soil hyphal length density in flat (a) and slope sites (b), mycorrhizal root intensity in flat (c) and slope sites (d) and arbuscule intensity in flat (e) and slope sites (f) in response to grazing intensities at three seasonal time points. Solid and hollow circles indicate mean and individual observations at each grazing intensity, respectively. Lines represent regressions from linear mixed-effects models, with solid and dashed lines indicating significant ( $P < 0.05$ ) and non-significant ( $P > 0.05$ ) relationships, respectively.



**Figure 2.** Estimated coefficients from general linear mixed models applied to (a) soil hyphal length density, (b) mycorrhizal root intensity and (c) arbuscule intensity in response to season in the flat and slope sites. Asterisks represent significance levels obtained from the model results,  $p < .001$ , "\*\*\*\*",  $p < .01$ , "\*\*\*",  $p < .05$ , "\*\*", NS: non-significant.

Variable	Model No	GI	Tp Slope	Season Summer	Season Autumn	GI: Tp Slope	GI: Season Summer	GI: Season Autumn	Tp Slope: Season Summer	Tp Slope: Season Autumn	GI: Tp Slope: Season Summer	GI: Tp Slope: Season Autumn	AICc
Soil hyphal length	1	-0.23±0.06 <b>(0.005)</b>	4.09±0.51 <b>(0.000)</b>	1.70±0.45 <b>(0.004)</b>	3.83± 0.36 <b>(0.000)</b>	-0.25±0.08 <b>(0.003)</b>	-0.04±0.08 (0.613)	-0.20±0.08 <b>(0.003)</b>	-3.59±0.75 <b>(0.000)</b>	-5.749±0.54 <b>(0.000)</b>	0.07± 0.14 (0.594)	0.14± 0.09 (0.257)	555.93
density (m/g)	2	-0.27±0.05 <b>(0.000)</b>	3.76±0.31 <b>(0.000)</b>	1.54±0.33 <b>(0.000)</b>	3.51±0.29 <b>(0.000)</b>	-0.18±0.04 <b>(0.000)</b>	-0.01±0.05 (0.906)	-0.13±0.05 <b>(0.007)</b>	-3.26±0.41 <b>(0.000)</b>	-5.11±0.29 <b>(0.000)</b>	-	-	553.75
Mycorrhizal root intensity	1	-0.91±0.43 (0.057)	9.56±3.29 <b>(0.004)</b>	-6.43±3.76 (0.052)	2.72±3.17 (0.391)	-1.12±0.61 (0.069)	0.87±0.61 (0.225)	0.27±0.59 (0.703)	-4.77±4.83 (0.324)	-15.53±5.37 <b>(0.001)</b>	0.99±0.89 (0.264)	0.84±0.83 (0.406)	1454.76
	2	-1.46±0.31 <b>(0.001)</b>	4.55±1.85 <b>(0.015)</b>	-8.68±2.63 <b>(0.001)</b>	0.84±2.56 (0.743)	-	1.37±0.41 <b>(0.001)</b>	0.68±0.41 (0.098)	-0.28±2.75 (0.920)	-11.77±2.48 <b>(0.000)</b>	-	-	1451.67
Arbuscule intensity	1	-0.42±0.15 <b>(0.028)</b>	1.74±1.14 (0.130)	-1.19±1.14 (0.359)	0.02±1.13 (0.989)	-0.15±0.21 (0.482)	0.27±0.21 (0.210)	0.11±0.21 (0.609)	0.88±1.63 (0.588)	-4.212±1.60 <b>(0.009)</b>	0.09±0.30 (0.745)	0.34±0.33 (0.257)	1019.72
	2	-0.49± 0.11 <b>(0.001)</b>	1.06± 0.63 (0.092)	-1.42± 0.91 (0.122)	-0.74± 0.91 (0.416)	-	0.31± 0.15 <b>(0.034)</b>	0.28± 0.15 (0.063)	1.32± 0.90 (0.143)	-2.70± 0.88 <b>(0.003)</b>	-	-	1014.33
Available phosphorus (mg/kg)	1	0.31± 0.08 <b>(0.004)</b>	-0.12± 0.64 (0.854)	0.63± 0.64 (0.327)	0.95± 0.55 (0.085)	-0.36± 0.12 <b>(0.003)</b>	-0.25± 0.12 <b>(0.037)</b>	-0.23± 0.10 <b>(0.026)</b>	-1.67± 1.03 (0.105)	-3.31± 0.77 <b>(0.000)</b>	0.28± 0.19 (0.140)	0.71± 0.14 <b>(0.000)</b>	725.43
Available nitrogen (mg/kg)	1	0.92±0.44 (0.059)	-6.23±3.38 (0.067)	10.28±3.38 <b>(0.003)</b>	2.06±2.82 (0.465)	-0.60±0.63 (0.341)	-1.48±0.63 <b>(0.019)</b>	-1.65±0.52 <b>(0.002)</b>	-4.29±5.46 (0.433)	-1.40±3.99 (0.726)	1.16±1.01 (0.253)	1.35±0.74 <b>(0.070)</b>	1415.57
	2	0.63±0.31 (0.068)	-8.91±1.88 <b>(0.000)</b>	7.67±2.51 <b>(0.003)</b>	-0.96±2.29 (0.676)	-	-0.90±0.37 <b>(0.016)</b>	-0.98±0.37 <b>(0.009)</b>	0.92±3.02 (0.761)	4.66±2.23 <b>(0.038)</b>	-	-	1413.36
Organic carbon (%)	1	-0.06± 0.03 <b>(0.052)</b>	-0.57± 0.22 <b>(0.011)</b>	-0.69± 0.22 <b>(0.002)</b>	-0.20± 0.14 (0.161)	0.049± 0.04 (0.236)	0.07± 0.04 (0.108)	-0.01± 0.03 (0.783)	1.3± 0.39 <b>(0.001)</b>	0.03± 0.20 (0.904)	-0.15± 0.07 <b>(0.038)</b>	0.04± 0.04 (0.279)	175.81
	2	-0.04±0.02 (0.094)	-0.35±0.14 <b>(0.012)</b>	-0.40±0.14 <b>(0.004)</b>	-0.24±0.08 <b>(0.004)</b>	-	-	-	0.62±0.25 <b>(0.016)</b>	0.21±0.11 (0.066)	-	-	173.25

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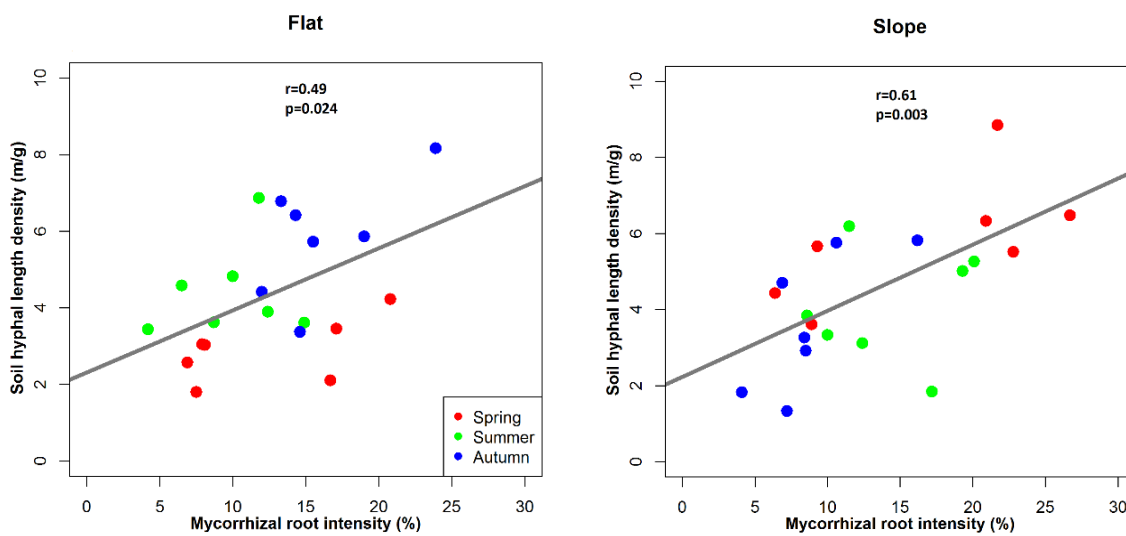
583 **Table-1 Linear mixed-effects models of the effects of grazing intensity (GI), topography (Tp) and season on AM fungi and soil variables.** Mixed-  
584 effects models were applied to nested (multi levels) data. The data are nested in the sense that samples were taken from two topographic locations  
585 (flat and slope) and in each topography from seven sites (called “plot” hereafter) representing seven levels of grazing intensity. In each plot,  
586 sampling was conducted at three seasons (spring, summer and autumn). The full model (model No. 1) and the best model selected according to  
587 Akaike’s information criteria (AIC) (model No. 2) are presented. Topography flat and season spring are reference groups in data presented here.  
588 Bold numbers represent the significant relationships (p<0.05).

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**Figure 3.** Pearson correlation coefficients ( $r$ ) between soil hyphal length density and mycorrhizal root colonization.

## Appendix A. Supplementary data

### SI-1 Site description

This experiment was set up at the Sino-German grazing experimental site which is located in the Xilin River Basin of Inner Mongolia, China (longitude 116° 42' E; latitude 43° 38' N). The experiment site was established in 2005 and is ca. 128 ha, with elevations of 1,200 m to 1,280 m asl. The area has a semi-arid, continental climate with a mean annual temperature of 0.9°C (1982–2010) and mean annual precipitation of 329 mm (1982–2010) with more than 70 % of the annual precipitation falling as rain during the growing season from April to September (Wan et al. 2011).

Two dominant plant species, *Leymus chinensis* (Trin.) Tzvel. and perennial bunchgrass *Stipa grandis* P. Smirn. together account for more than 75% of the total above-ground biomass (Li et al. 2017). Other species that commonly appeared in our experimental site are *Cleistogenes squarrosa* (Trin.) Keng, *Agropyron cristatum* (L.) Gaertn., *Koeleria cristata* (L.) Pers., *Achnatherum sibiricum* (L.) Keng, *Carex korshinskyi* Kom., *Potentilla acaulis* L., *Allium bidentatum* Fisch. ex Prokh., *Allium tenuissimum* L., *Chenopodium aristatum* L., *Salsola collina* Pall., and *Chenopodium glaucum* L. (Schönbach et al. 2011, Li et al. 2017). The total vegetation cover is about 30-40% in normal years and may reach 60-70% in wet years.

The pastures are generally grazed by sheep and goats. The major soil type is calcic chernozem (IUSS Working Group 2006), developed from aeolian sediments deposited on a Pleistocene basalt plateau, with mainly a fine-sand loess texture. The soils are defined by a dark Ah horizon followed by an Ach horizon. The carbonate-free Ah horizon thickness differs from 20 to more than 100 cm where *L. chinensis* is dominant and from 5 to 45 cm where *S. grandis* sites is abundant. The Ach horizon containing secondary calcium carbonate nodules is located bellow the Ah horizon (Wiesmeier et al. 2009).

## SI-2 Hyphal length density (HLD) measurement

Before sub-sampling and hyphal extraction, soil samples were passed through a 2.00 mm sieve to remove large particles and roots. A modified membrane filter technique (based primarily on Jakobsen et al. (1992) and Boddington et al. (1999)) was used for soil hyphal extraction as follows: from each soil core, two 5 g sub-samples were taken for external mycorrhizal hyphae extraction (140 samples in total). 5 g soil in 500 ml of deionized water (dH<sub>2</sub>O) was stirred at full speed with a magnetic stirrer for 2 min in a beaker, and afterward was poured through a 0.5 mm wire sieve to collect large particles. The solution was agitated in 500 ml by stirring for 10 s with a glass rod and allowed to settle for 10 s and then decanted through a 45 µm sieve. Agitation was repeated three times to ensure most of the hyphae were obtained. Finally, the material on the 45 µm sieve was rinsed into a beaker using 250 ml dH<sub>2</sub>O, then placed in a filter cylinder fitted with a 0.45 µm nitrocellulose membrane filter. The samples were left to drain under vacuum. With the vacuum on, cylinders were removed and a few drops of Trypan blue staining solution (lactic acid: glycerol: Trypan blue (5%v/w); 1:2:1) was carefully added to the filter containing the extracted material. The membrane filters, containing the extracted hyphae, were removed after drying and rinsed in lactic acid: glycerol: dH<sub>2</sub>O (1:2:1) and then cut into two pieces and transferred to microscope slides. Assessment of hyphal length per filter was carried out by the gridline intercept method for a minimum 60 fields of view for each filter paper at × 100 magnification (using a 10 × 10 grid of 5 mm length formed by 11 horizontal and 11 vertical lines intercrossed perpendicularly). The hyphae that were angular, aseptate, and 1.0–13.4 µm in diameter were deemed to be of AMF origin (Boddington et al. 1999, Shen et al. 2016), and only those were considered for the measurements. The developed modified GIM (Gridline Intersect Method) equations based on (Tennant 1975) were used for calculating the total length of hyphae (mm) (Shen et al. 2016) as follows:

$$L = \frac{\text{Total number of intersections} \times \frac{11}{14} \times g \times Af}{Ag \times N}$$

Where:

$\frac{11}{14}$  is a constant

"g" is the grid unit

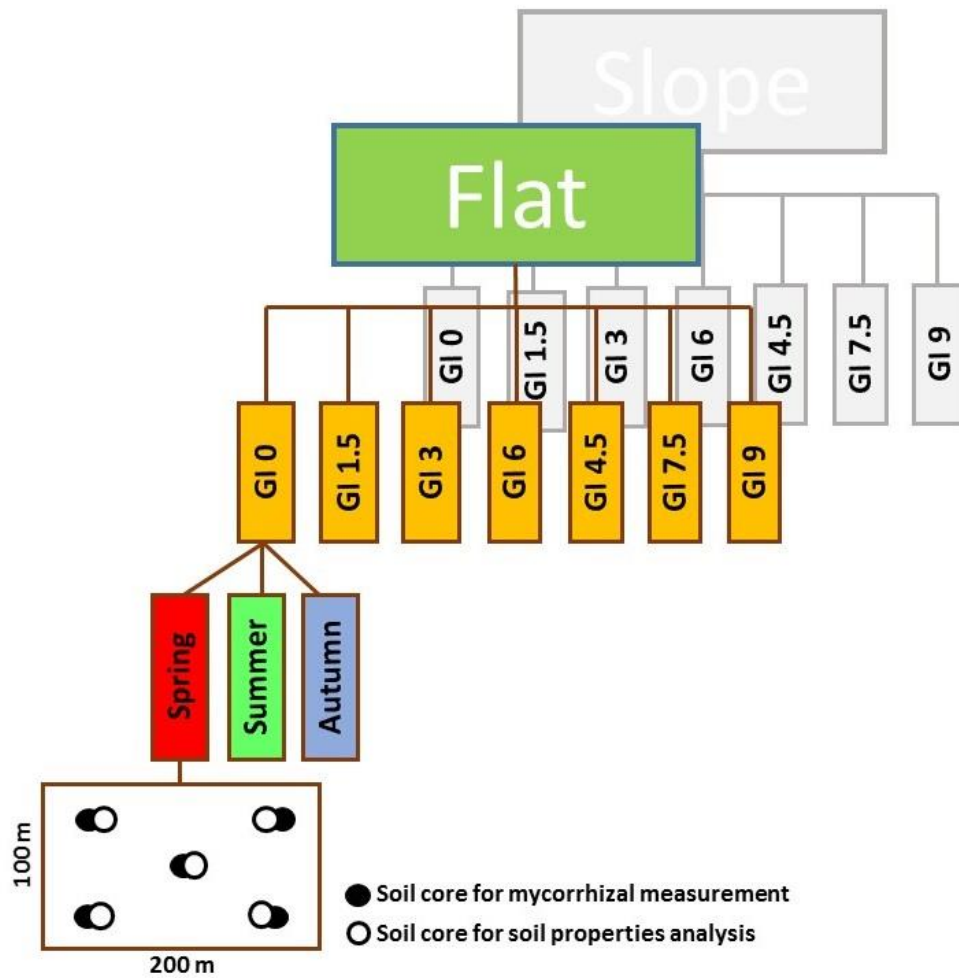
Af is the area of the filter

Ag is the area of the grid unit

N is the number of fields of view on each filter

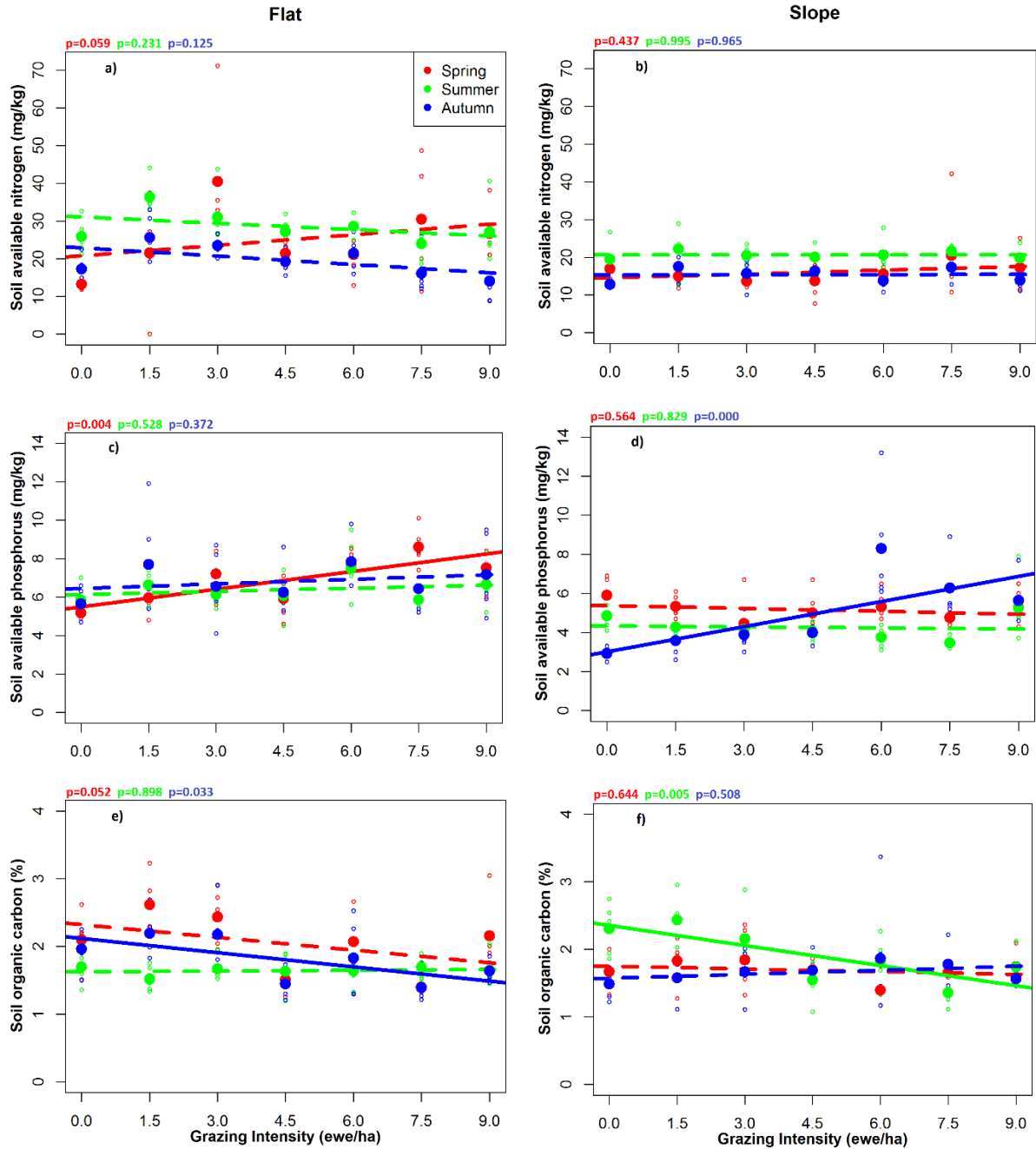
"Total number of intersections" is the count of the number of intersections across vertical and horizontal lines for each filter

### SI-3 Schematic illustration of experimental design



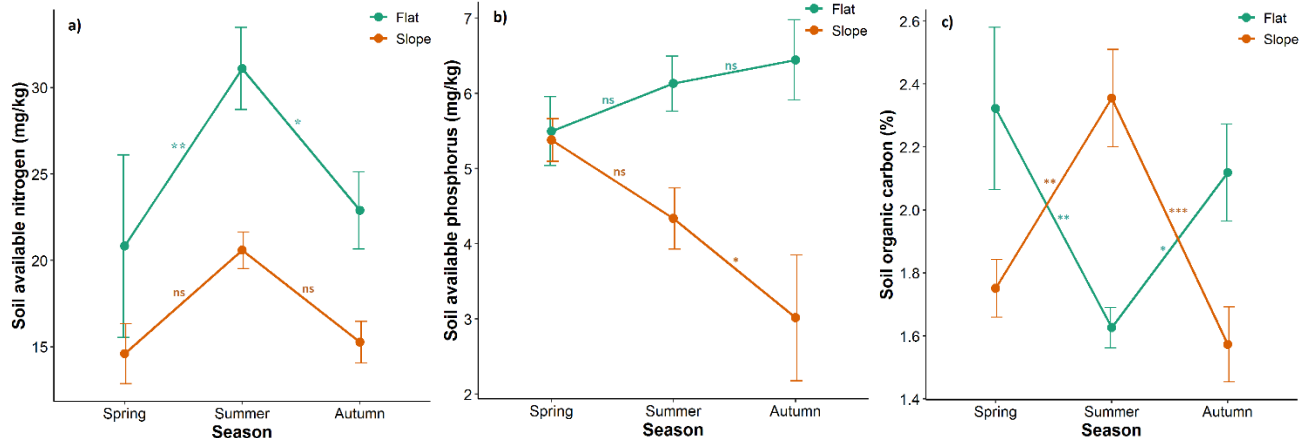
**Figure S1.** Schematic illustration of experimental design, grazing plots and sample collection. Sampling was conducted at two locations with contrasting topography; flat and slope. At each topographical location, samples were collected from seven plots, each with a different grazing intensity (GI); 0 (no grazing), 1.5 (very light), 3 (light), 4.5 (light-moderate), 6 (moderate), 7.5 (heavy) and 9 (overgrazing). Each grazing plot has an area of 2 ha. Double soil core samples were taken at each sampling point, one for mycorrhizal measurement and assessment and one for soil properties analyses. Sampling was repeated in three seasons; spring, summer and autumn.

# **SI-4 Seasonal variations in soil available nitrogen, phosphorus and soil organic carbon along a grazing gradient at two topographical locations.**



**Figure S2.** Soil available nitrogen in topography flat (a) and slope (b), soil available phosphorus in topography flat (c) and slope (d) and soil organic carbon in topography flat (e) and slope (f) in response to grazing gradient at three seasons. Solid and hollow circles indicate means and individual observations at each grazing intensity, respectively. Lines represent regressions from linear mixed-effects models, with solid and dashed lines indicating significant ( $P < 0.05$ ) and non-significant ( $P > 0.05$ ) relationships, respectively.

1 **SI-5 variation in soil resource availability over the course of growing season**



3 **Figure S3.** Estimated coefficients from general linear mixed models applied on available nitrogen (a), available  
4 phosphorus (b), and organic carbon (c) in response to season in flat and slope sites. Asterisks represent significance  
5 level obtained from the model results,  $p < .001$ , "\*\*\*\*",  $p < .01$ , "\*\*\*",  $p < .05$ , "\*\*",  $p < .1$ , "\*", NS: non-significant.

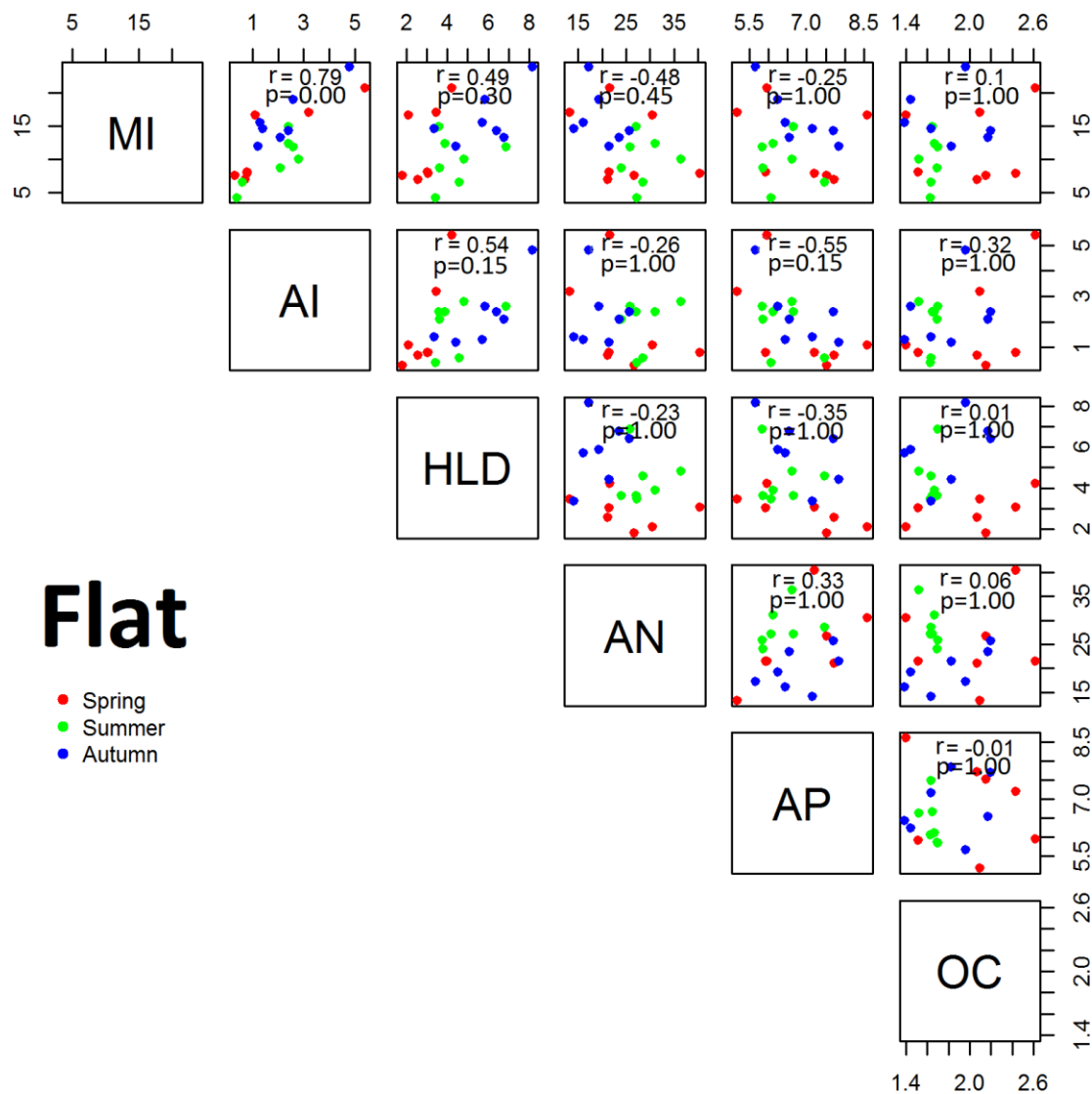
8 **SI-6 Linear regression model of the relationship between AMF fungal and soil resource**  
9 **variables**

11 **Table S1- Linear regression model of the relationship between AMF fungal and soil resource variables**

Soil variables	Topography	Soil hyphal length density (m/g)	Mycorrhizal Intensity (%)	Arbuscule Intensity (%)
Organic Carbon (%)	Flat	0.03±1.12 (0.978)	1.49±3.37(0.664)	1.25±0.85(0.157)
	Slope	1.17 ± 1.55(0.462)	7.35±5.26(0.179)	5.22±1.84( <b>0.011</b> )*
Available Nitrogen (mg/kg)	Flat	-0.06± 0.06 (0.317)	-0.36±0.15( <b>0.029</b> )*	-0.05±0.044(0.255)
	Slope	-0.10±0.14 (0.483)	-0.28±0.50(0.579)	0.27±0.19(0.167)
Available Phosphorus (mg/kg)	Flat	-0.67±0.41(0.118)	-1.45±1.28 (0.274)	-0.84±0.29 ( <b>0.01</b> )**
	Slope	-0.24±0.35(0.509)	-0.66±1.24(0.599)	-0.29±0.49(0.554)

12 Values indicate slope coefficients ± SE (p-value) extracted from linear regression models between AMF variables as dependent variables  
13 and soil resource variables as independent variables. Boldface indicates significant relationships ( $p < 0.05$ ) and asterisks represent  
14 significance level obtained from the model results,  $p < .001$ , "\*\*\*\*",  $p < .01$ , "\*\*\*",  $p < .05$ , "\*\*",  $p < .1$ , "\*", NS: non-significant.

21 **SI-7 Correlogram of AMF and soil resource variables in the flat site**



22

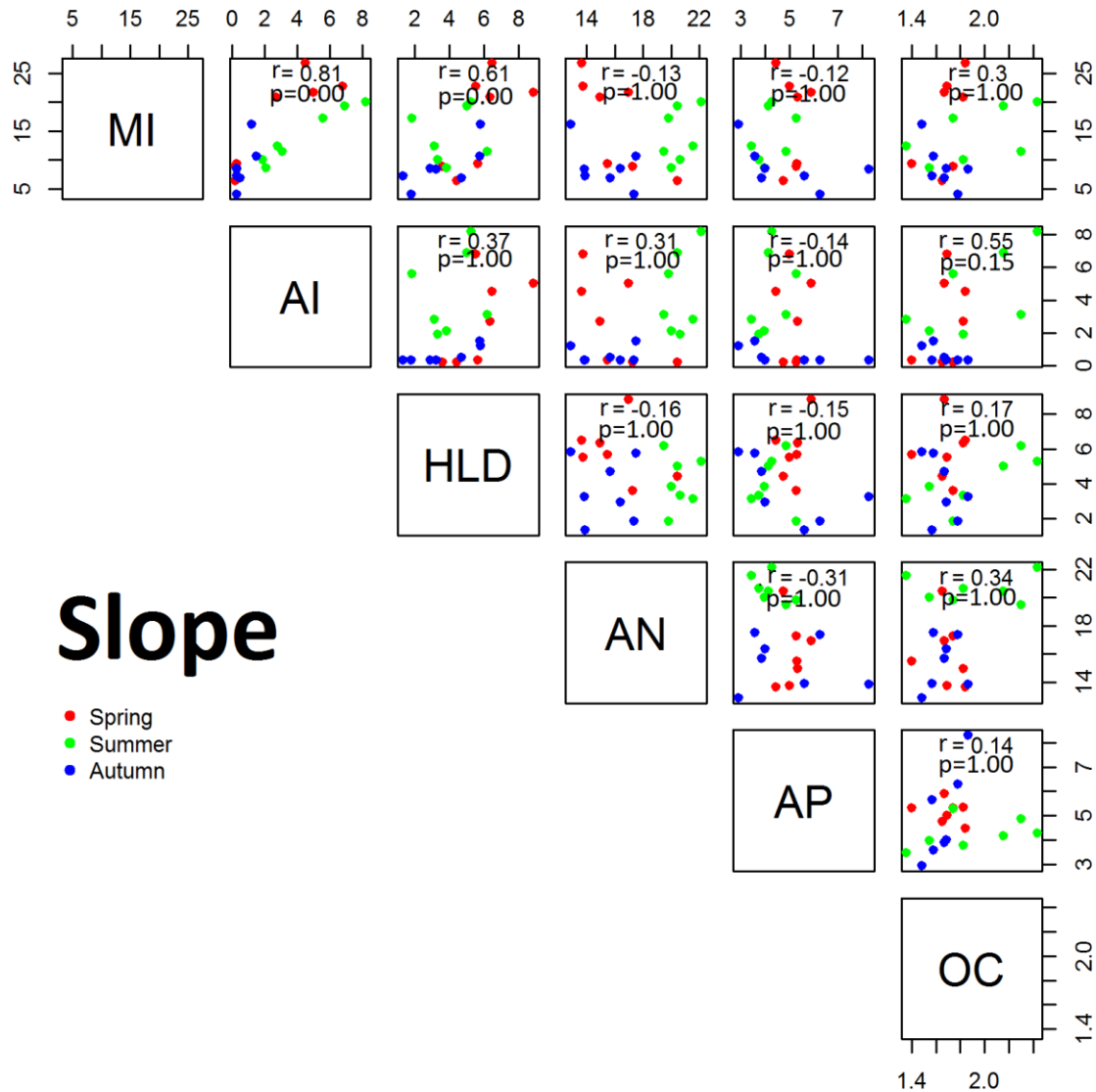
23 **Figure S4.** Correlogram of AMF and soil resource variables in the flat site. Abbreviation: MI: mycorrhizal  
24 intensity (%), AI: Arbuscule intensity (%), HLD: soil hyphal length density (m/g), AN: soil available nitrogen  
25 (mg/kg), AP: soil available phosphorus (mg/kg) and OC: soil organic carbon (%). Bonferroni correction has  
26 been applied to calculate the adjusted p-values.

27

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SI-8 Correlogram of AMF and soil resource variables in slope site



**Figure S5.** Correlogram of AMF and soil resource variables in slope site. Abbreviation: MI: mycorrhizal intensity (%), AI: Arbuscule intensity (%), HLD: soil hyphal length density (m/g), AN: soil available nitrogen (mg/kg), AP: soil available phosphorus (mg/kg) and OC: soil organic carbon (%). Bonferroni correction has been applied to calculate the adjusted p-values.



**Table S2-The best-fit model of grazing intensity effects on AMF and soil recourse variables based on AIC.**  $\Delta$ AIC is the difference in AIC between the full model and the best-fit model.

	The best-fitting model	$\Delta$ AIC
Soil hyphal length density (m/g)	GI+ Tp + Season+ GI×Tp + GI×Season + Tp×Season	2.180
Mcorrhizal root intensity (%)	GI+ Tp + Season+ GI×Season + Tp×Season	3.094
Arbuscule intensity (%)	GI+ Tp + Season+ GI×Season + Tp×Season	5.374
Available phosphorus (mg/kg)	GI+ Tp + Season+ GI×Tp + GI×Season + GI×Season + GI×Tp×Season	0.000
Available nitrogen (mg/kg)	GI+ Tp + Season+ GI×Season + GI×Season	2.210
Organic carbon (%)	GI+ Tp + Season+ GI×Tp + GI×Season	2.558

Note: full model: GI+ Tp + Season+ GI×Tp + GI×Season + GI×Season + GI×Tp×Season. Tp: topography and GI: grazing intensity. The best fitting model is the model with lowest AIC.

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